

**Some notes on diffusion of radio-  
nuclides through compacted clays**

Trygve E Eriksen

The Royal Institute of Technology, Department of Nuclear  
Chemistry, Stockholm, Sweden

May 1989

SOME NOTES ON DIFFUSION OF RADIONUCLIDES THROUGH  
COMPACTED CLAYS

Trygve E Eriksen

The Royal Institute of Technology, Department of  
Nuclear Chemistry, Stockholm, Sweden

May 1989

This report concerns a study which was conducted for SKB. The conclusions and viewpoints presented in the report are those of the author(s) and do not necessarily coincide with those of the client.

Information on SKB technical reports from 1977-1978 (TR 121), 1979 (TR 79-28), 1980 (TR 80-26), 1981 (TR 81-17), 1982 (TR 82-28), 1983 (TR 83-77), 1984 (TR 85-01), 1985 (TR 85-20), 1986 (TR 86-31), 1987 (TR 87-33) and 1988 (TR 88-32) is available through SKB.

Some notes on diffusion of radionuclides

through compacted clays.

Trygve E Eriksen

The Royal Institute of Technology,

Department of Nuclear Chemistry,

S-100 44 Stockholm 70

Abstract.

The apparent diffusivities of some simple cations i.e.  $\text{Sr}^{2+}$ ,  $\text{Cs}^+$  in water saturated bentonite indicates that diffusion takes place both in the pore solution and within the solvation sheath of the exchangeable cations. Data from some earlier diffusion experiments have been re-evaluated and the results suggest that diffusion within the solvation sheath of the exchangeable cations is the dominating mechanism for  $\text{Sr}^{2+}$  and  $\text{Cs}^+$ .

Content.

	page
Abstract	i
Table of contents	ii
1. Introduction	1
2. Components of diffusion coefficients	2
3. Clay-water interaction	4
4. Experimental techniques	5
4.1 Through diffusion	5
4.2 Back to back experiments	7
4.3 Material	9
5.1 Anion diffusion	10
5.2 Cation diffusion	11
References	16

## 1. Introduction.

To make possible the assessment of long term storage of high level nuclear wastes in underground repositories understanding of the migration mechanisms in porous water containing materials is of great importance.

A clay backfill is generally assumed to create a first barrier to the transport of radionuclides when the canisters fail and diffusion through compacted water saturated bentonite clays has been the subject of several studies (1-12).

The experimental techniques used and the results obtained have earlier been discussed in some detail by Eriksen and Jacobsson (13). Methods of measuring diffusion properties of porous media have also recently been discussed by Lever (14).

The transport of solutes is of great importance for many soil processes and the diffusion of ions and uncharged solutes through soil and clays have recently been comprehensively reviewed by Nye (15).

Applying Fick's first law  $F = -D \cdot \partial c / \partial x$ , where  $F$  is the flux,  $D$  the diffusion coefficient and  $\partial c / \partial x$  the concentration gradient, to a complex medium such as water saturated bentonite may be regarded as giving an operational definition of the diffusion coefficient. To explain the results obtained at varying experimental conditions a thorough understanding of the clay/water system is, therefore, required.

Some of the models used to describe the diffusion in compacted bentonite will be discussed below.

## 2. Components of diffusion coefficients.

In the overall movement of solutes through a water containing porous medium both solid and water play a part. If the diffusion is assumed to take place in the pore water only the flux per unit area of of the porous medium is linked to the concentration gradient in the pore water by the intrinsic diffusion coefficient for the pore water  $D_{ip}$

$$F = -D_{ip} \cdot \nabla C_p \quad /1/$$

If it is assumed that the solute is sorbed on the surface of the porous medium by a sorption reaction that is instantaneous and reversible the diffusion in one dimension is described by the equation

$$\theta \cdot \delta C_p / \delta t = D_{ip} \cdot \delta^2 C_p / \delta x^2 - \delta C_s / \delta t \quad /2/$$

where  $C_p$  is the concentration of solute in pore water,  $C_s$  is the concentration of sorbed solute and  $\theta$  is the volume fraction of the pore water.

The quantity of solute sorbed is frequently measured per unit mass of porous medium and the ratio of the concentration of solute bound to the solid phase relative to the concentration in solution is given by the distribution coefficient  $K_d$ . Based on the assumptions given above the concentration of sorbed solute is related to the solute concentration in the pore water by the density of the porous medium and the distribution coefficient.

$$C_s = (K_d \cdot \rho) C_p \quad /3/$$

and equation /2/ can be rewritten as

$$\alpha \cdot \delta C_p / \delta t = D_{ip} \cdot \delta^2 C_p / \delta x^2 \quad /4/$$

where  $\alpha = \theta + K_d \cdot \rho$  is the capacity factor of the porous medium. The total quantity of solute per unit volume of the porous medium is thus  $\alpha \cdot C_p$  and the total flux per unit area is  $D_{ip} \cdot \delta C_p / \delta x$

The apparent diffusivity is a measure of the distance that the solute has diffused and the appropriate diffusion coefficient is

$$D_a = D_{ip} / \alpha \quad /5/$$

For weakly or non sorbing solutes

$$D_a = D_{ip} / \theta = D_p \quad /6/$$

where  $D_p$  is the pore-water diffusion coefficient.

The experimental data on cation diffusion in compacted clays can not be easily accommodated by this pore diffusion model and several researchers (16,17) have recently argued for the inclusion of an additional surface flux  $F_s$  that arises from the



diffusion of the solute sorbed on the surface of the porous solid.

The total flux is in this case given by

$$F = -D_{ip} \cdot \nabla C_p - D_s \cdot \nabla C_s \quad /7/$$

where  $D_s$  is the surface diffusion coefficient and equations 4 and 5 are modified to read

$$\alpha \cdot \delta C_p / \delta t = D_i \cdot \delta^2 C_p / \delta x^2 \quad /8/$$

where  $D_i = D_{ip} + K_d \cdot \rho \cdot D_s$  is the total intrinsic diffusion coefficient and

$$D_a = (D_{ip} + K_d \cdot \rho \cdot D_s) / \alpha \quad /9/$$

### 3. Clay- water interaction.

The clay used in many of the studies cited is the American Colloid Co type Mx-80 granulated Na bentonite, containing about 75% montmorillonite.

The mechanism of hydration and the structure of smectite-adsorbed water have recently been reviewed (18-19). It seems clear that the first stage in water adsorption involves solvation of the exchangeable cations (mainly in the interlamellar positions), the degree of solvation depending on the ability of the cations to solvate themselves. The ordered arrangement thus obtained for the first 1-3 first layers of water in contact with the surface remains unaltered even on addition of further water. As the water content is increased

water is adsorbed on external surfaces and in micropores and the behaviour of the hydrated cations becomes gradually more similar to that of cations in aqueous solution. At a given water content the water adsorbed by smectites can thus be found in two different states. A general relationship between bulk density and content of internal water in percent of the total water volume in smectite clay is given in a recent paper by Pusch (20).

#### 4. Experimental techniques.

Several different techniques have been used for measuring diffusion properties of porous media. In this section we give a short description of some tracer techniques used to study diffusion in highly compacted bentonites.

##### 4.1 Through diffusion.

A compacted water saturated bentonite is sandwiched between two reservoirs (Fig 1).

The tracer is added to one of the reservoirs and the temporal change in the tracer concentration in the other reservoir monitored. The total amount of tracer  $Q(t) = C_2(t)V_2$ , accumulated in the initially tracer free reservoir is given by

$$Q/A \cdot l \cdot C_1 = D_i t/l^2 - \alpha/6 - (2\alpha/\pi) \sum_{n=1}^{\infty} (-1)^n/n^2 \exp -D_i n^2 \pi^2 t/l^2 \quad /10/$$

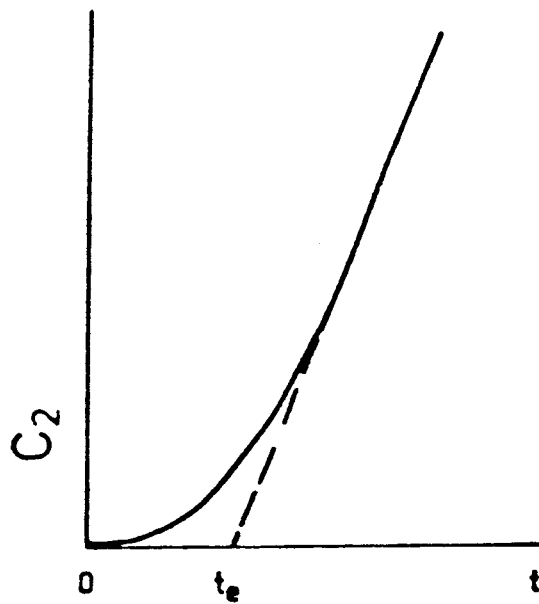
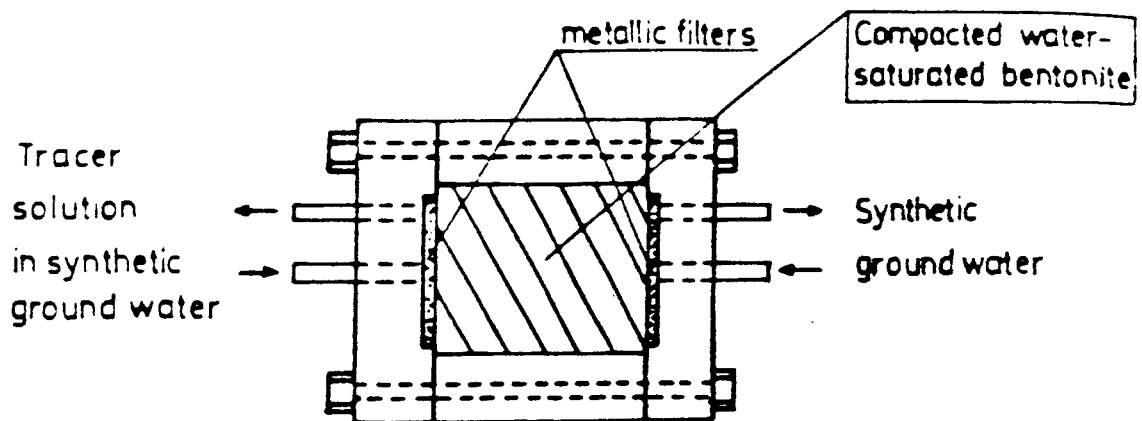


Fig 1

With increasing  $t$  the exponential term rapidly decreases leaving the asymptotic solution

$$C_2 = (AC_1/V_2l)(D_1t - \alpha l^2/6)$$

/11/

where  $D_i$  is the intrinsic diffusion coefficient,  $C_1$  the inlet concentration,  $C_2$  the concentration in the outlet reservoir with volume  $V_2$ ,  $A$  the cross-sectional area of the clay sample,  $l$  the sample thickness and  $\alpha$  the capacity factor.

The asymptotic solution thus gives a straight line with slope  $D_i(AC_1/V_2l)$  and an intercept on the  $t$ -axis at  $t_e = \alpha l^2/6D_i$ . The capacity factor can be obtained from the intercept of the extrapolated asymptote with the  $C_2$ -axis. i.e.  $C_2(0) = (AC_1l/6V_2) \cdot \alpha$ .

#### 4.2 Back to back experiments.

A water saturated compacted clay sample is pressed in contact with a corresponding clay sample containing the tracer.

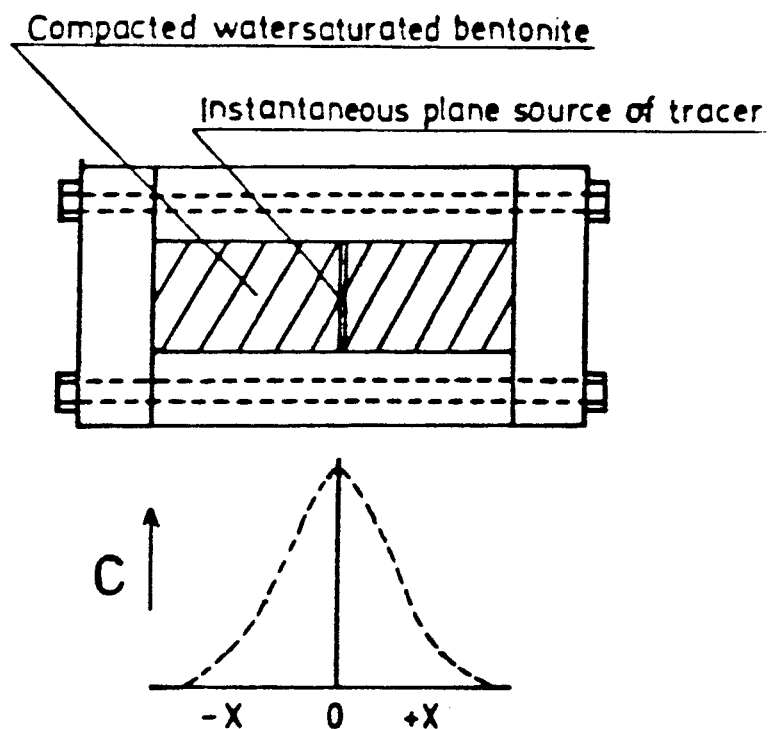


Fig 2

If the tracer is applied to a thin layer and the total amount of tracer is  $M$  then the tracer concentration at a distance  $x$  from the initial source at time  $t$  is

$$C(t) = [M/2A(\pi D_a t)^{1/2}] \cdot \exp(-x^2/4D_a t) \quad /12/$$

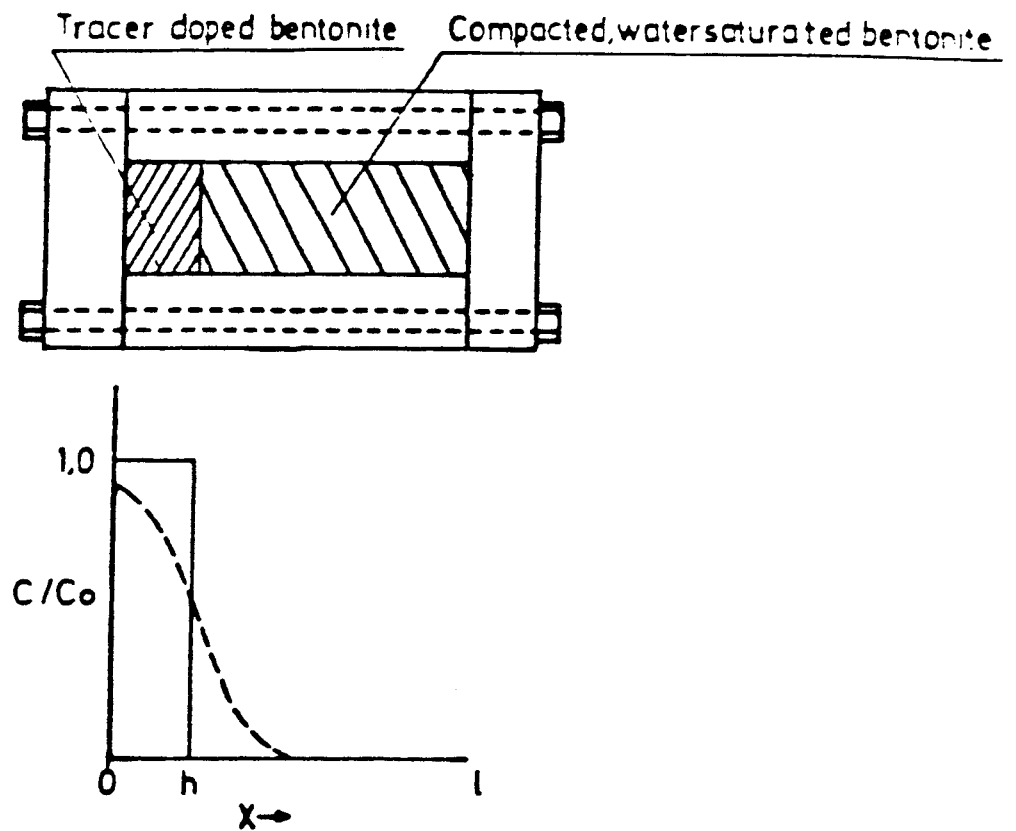


Fig 3

If one half is uniformly loaded with the tracer i.e extended initial distribution the tracer distribution is given by the equation

$$C/C_0 = 1/2 \operatorname{erf}[(l-x/h)/(2/h) \cdot (D_a t)^{1/2}] + 1/2 \operatorname{erf}[(l+x/h)/(2/h) \cdot (D_a t)^{1/2}] \quad /13/$$

where  $C_0$  is the initial tracer concentration in the loaded sample of thickness  $h$  and  $l$  the overall thickness of the clay samples.

In both cases only the apparent diffusion coefficient can be calculated directly from the diffusion data and the capacity factor must be determined in separate sorption experiments.

#### 4.3 Material.

The clay used in the experiments was the American Colloid Co type MX-80 granulated Na bentonite. The bentonite was dried at 105°C, compacted and thereafter water saturated. The experimental conditions are given in table 1.

Table 1.

Properties of diffusing medium.

Dry density $g \cdot cm^{-3}$	Bulk density $g \cdot cm^{-3}$	$V_w/V_w+V_s$ %	Water ratio %
1.8	2.12	32	17.7

### 5.1 Anion diffusion.

The experiments were carried out using the through diffusion technique and the experimental results are given in table 2.

Table 2.

Anion diffusion in compacted Mx-80. (Bulk density  $2.12 \text{ g}\cdot\text{cm}^{-3}$ )

Ion	$D_a \cdot 10^{12}$ $\text{m}^2\cdot\text{sec}^{-1}$	$D_i \cdot 10^{14}$ $\text{m}^2\cdot\text{sec}^{-1}$	$\alpha$	$V_w/V_w+V_s$	$\alpha/(V_w/V_w+V_s)$
$\text{Cl}^-$	92	31	0.0034	0.32	0.011
$\text{I}^-$	35	21	0.0060	0.32	0.019
$\text{HS}^-$	4.8	1.14	0.0024	0.32	0.008
$\text{AQS}^{-*}$	8.2	0.23	0.0003	0.32	0.001
$\text{LS}^{-\S}$	5.6	0.075	0.00013	0.32	0.0004

\* AQS = Antraquinone sulfonate

\\$ LS = Lignosulfonate (Mw 30 000)

Assuming no sorption the relation between the pore diffusion coefficient  $D_p$  and the free water diffusion coefficient  $D_w$  can be written  $D_p = f \cdot D_w$  where  $f$  is an impedance factor.

In the case of uncharged solutes the impedance factor takes account primarily of tortuosity ( $\tau^2$ ) and constrictivity ( $\delta$ ) and can be written  $f = \delta/\tau^2$ . In the case of anion diffusion the impedance factor also allows for Donnan exclusion.

Taking the free water diffusion coefficient of  $\text{Cl}^-$  to be  $2.03 \cdot 10^{-9} \text{ m}^2 \text{ s}^{-1}$  the impedance factor is 0.045, indicating marked constrictivity and/or tortuosity. The capacity factor is also very low compared to the volumetric water ratio. It is thus very clear that very little, if any, interlamellar anion diffusion takes place. If the anion "diffusion porosity" is assumed to represent the external pore-water (as described by Pusch) the  $\text{Cl}^-$ ,  $\text{I}^-$  and  $\text{HS}^-$  data indicate that with the exception of a few percent, the water is present as interlamellar water.

### 5.2 Cation diffusion.

Diffusion coefficients for  $\text{Sr}^{2+}$  and  $\text{Cs}^+$  are given in table 3.

Table 3.

$\text{Sr}^{2+}$  and  $\text{Cs}^+$  diffusion in compacted Na- bentonite.  
(Bulk density  $2.1 \text{ g} \cdot \text{cm}^{-3}$ ).

Ion	Ref	$D_a \cdot 10^{12}$ $\text{m}^2 \cdot \text{s}^{-1}$	$V_w/V_w+V_s$	$K_d$ $\text{cm}^3 \cdot \text{g}^{-1}$	$D_w \cdot 10^9$ $\text{m}^2 \cdot \text{s}^{-1}$
$\text{Sr}^{2+}$	6	15+/-2	0.32	604	0.78
	5.9	1.5-12		2900	
$\text{Cs}^+$	6	2.5+/-0.5	0.32	645	2.11
	9	1.4-2		1400	
	10	5.0			



According to the pore diffusion model, assuming no diffusion of sorbed exchangeable ions, the apparent diffusivity is described by  $D_a = D_{ip}/\alpha$  where  $\alpha = \theta + K_d \cdot \rho$ .

The impedance factor  $f = \delta/\tau^2$  is not known, but as we have not used oriented samples it is not likely to be higher than 0.5. Based on the free water diffusivities and  $K_d$  values given in table 3, the upper limits for  $D_a$  are calculated to be  $0.34 \cdot 10^{-12}$  and  $0.9 \cdot 10^{-12} \text{ cm}^2 \cdot \text{s}^{-1}$  for  $\text{Sr}^{2+}$  and  $\text{Cs}^+$  respectively, i.e. lower than the experimentally measured apparent diffusivities. The experimental data can thus not be accommodated by the pore diffusion model.

In a study on  $\text{Sr}^{2+}$ ,  $\text{Cs}^+$  and  $\text{Co}^+$  diffusion in compacted bentonite Murinen and co-workers varied the the  $K_d$  values by changing the ionic strength of the solutions used for water saturation of the bentonite. The electrolyte used was NaCl and the highest concentration was  $0.6 \text{ mol} \cdot \text{dm}^{-3}$ . By fitting the data to a model with surface diffusion included they calculated  $\theta \cdot D_p$  and  $D_s$ . The surface diffusion was found to account for 70-90 percent of the total transport in their experiments.

The idea of separating the cation migration in clays into two components i.e. the flux through the pore solution alone and the flux of exchangeable cations on the solid is not a new one. The overall cation transport in suspensions with low clay content (21, 22) as well as on water saturated clay films (23) have been explained by a model based on the assumption of diffusion taking place both in a solution phase and in the adsorbed state. The application of this concept does, however, require knowledge of the volume fraction occupied by the pore

solution. It ought to be emphasized that the water associated with the sorbed exchangeable ions is not to be included.

Mott (24) measured the diffusion coefficients of  $\text{Na}^+$  and  $\text{Sr}^{2+}$  in oriented homoionic bentonite flakes with varying water content. He found the diffusion in dry clay to be very slow ( $D_a < 10^{-15} \text{ m}^2\text{s}^{-1}$ ) and to increase with several orders of magnitude on addition of 0.15-0.3 g water/g dry clay. This water content corresponds to 1-3 molecule layers of water.

A detailed study on diffusion of the exchangeable cations  $\text{Na}^+$ ,  $\text{Rb}^+$ ,  $\text{Cs}^+$  and  $\text{Ca}^{2+}$  in montmorillonite containing 0-0.25 g water/ g dry clay was carried out by Calvet (24). He found the mobility of  $\text{Na}^+$  to be much reduced on addition of  $\text{Ca}^{2+}$  to a homoionic clay and that the effect was enhanced at low water content. According to Calvet this effect is due to the higher water affinity of  $\text{Ca}^{2+}$  as compared to  $\text{Na}^+$ , i.e competition for water in the slightly hydrated clay.

In the highly compacted bentonite ( $\rho = 2.12 \text{ g}\cdot\text{cm}^3$ ) used in our experiments the water content corresponds to 2-3 molecule layers of water. Most of this water is associated with the exchangeable cations and is thereby situated in interlamellar positions. Pusch (20) estimates that 75-85 percent of the water is present as "internal" water at the bulk density used in our experiment.

If it is assumed that the impedance factor for the cation transport in the pore solution (i.e interlamellar water not included) is the same as for the anions  $\text{Cl}^-$  and  $\text{I}^-$  and that the interlamellar water is totally associated with the exchangeable cations the diffusion coefficients in pore solution and in the

solvated cation phase can be calculated. The results are tabulated in table 4.

Table 4.

Calculated diffusion coefficients for pore solution and solvated-cation phase.

Ion	$D_a \cdot 10^{12}$ $m^2 s^{-1}$	f	$D_w \cdot 10^9$ $m^2 s^{-1}$	$D_{ip} \cdot 10^{11}$ $m^2 s^{-1}$	$D_s \cdot 10^{12}$ $m^2 s^{-1}$
$Sr^{2+}$	15	0.045	0.78	3.5	$15^a$ $12-13.5^b$
$Cs^+$	2.5	0.045	2.1	9.5	$2.1^a$ $0.5-1.3^b$

a : based on pore solution porosity of 0.0045

b : based on pore solution porosity of 15-25% of total water volume fraction (according to Pusch, ref 20).

The diffusion coefficients, calculated in this way, are somewhat higher, but of the same order of magnitude as those obtained by Murinen et al (17).

Some objections can be raised against the use of solutions with varying concentrations of electrolytes and complexing agents to vary the  $K_D$ -value in diffusion experiments. At low water ratios a redistribution of the water on the clay surface probably takes place which, according to the work of Calvet (25), may influence the diffusivity of the exchangeable ions.

All the experimental data indicate that diffusion within the solvation sheath of the exchangeable ions is a dominating transport process in highly compacted bentonite with low water content.

References.

1. Neretnieks I, Skagius K  
Diffusionsmätningar i våt lera.  
KBS TR-87
2. Neretnieks I, Skagius K  
Diffusionsmätningar av metan och väte i våt lera.  
KBS TR 86
3. Allard B, Kipatsi H  
Mätning av diffusionshastigheter för silver i lera-  
sandblandning.  
KBS TR 53
4. Eriksen T E, Jacobsson A, Pusch R  
Ion diffusion through highly compacted bentonite.  
KBS TR 81-06
5. Torstenfelt B, Andersson K, Kipatsi H, Allard B,  
Olofsson U  
Diffusion measurements in compacted bentonite.  
Topp S (Ed): Scientific Basis for Nuclear Waste  
Management 1V, Elsevier 1982.
6. Eriksen T E, Jacobsson A  
Ion diffusion in compacted sodium and calcium bentonites.  
KBS TR 81-12.

7. Eriksen T E and Jacobsson A  
Diffusion of hydrogen,hydrogensulfide and large molecular weight anions in bentonite.  
KBS TR 81-17
  
8. Pusch R, Eriksen T E and Jacobsson A  
Ion/water migration phenomena in dense bentonites.  
Lutze W (Ed): Scientific Basis for Nuclear Waste Management V, Elsevier 1982 p.649
  
9. Torstenfelt B, Kipatsi H, Andersson K, Allard B, Olofsson U.  
Transport of actinides through a bentonite backfill.  
ibid p.659
  
10. Murinen A,Rantanen R,Ovaskinen R, Heinonen O J  
Diffusion measurements in concrete and compacted bentonite.  
Brookins D G (Ed): Scientific Basis for Nuclear Waste Management VI, Elsevier
  
11. Kipatsi H  
Sorption behaviour of long lived fission products and actinides in clay and rock.  
Dissertation, Univ. Gothenburgh 1983.

## 12. Neretnieks I

Diffusivities of some dissolved constituents in compacted wet bentonite clay-MX 80 and the impact on radionuclide migration in the buffer.

KBS TR 82-27

## 13. Eriksen T E, Jacobsson A

Diffusion in clay, experimental techniques and theoretical models.

KBS TR 84-05

## 14. Lever D A

Some notes on experiments measuring diffusion of sorbed nuclides through porous media.

AERE R 12 321 (1986)

## 15. Nye P H

Diffusion of ions and uncharged solutes in soils and soil clay.

Adv in Agronomy Vol 31, P 225-272. A.P 1979.

## 16. Rasmusson A, Neretnieks I

Surface migration in sorption processes.

KBS TR 83-37.

17. Murinen A, Rantanen J, Hiltunen P P  
Diffusion mechanisms of strontium, cesium and cobalt in compacted sodium bentonite.  
Werme L (Ed): Scientific Basis for Nuclear Waste Management IX, Material Research Society Symposia Proceedings Vol 50, p 617 (1985).
  
18. Fripiat J J, Stone W E E  
Water on the surface of clay minerals: orientation, diffusion and proton exchange.  
Rev. Chem. Liq. Vol 7, p 349-374 Gordon and Breach 1978.
  
19. Sposito G, Prost R  
Structure of water adsorbed on smectites  
Chem.Rev: 1982,86:6,553
  
20. Pusch R, Karnland O  
Aspects of the physical state of smectite- adsorbed water.  
SKB TR-86-27
  
21. Duffey J E, Landelout H G  
Self diffusion of anions in clay- gels  
J.Colloid. Interface Sci. 1975, 51, 278



22. Duffey J E, Landelout H G

Self diffusion of sodium on clay surfaces as influenced  
by two other alkali cations.

J. Colloid. Interface Sci. 1975, 52,340

23. Ellis J H, Bornhisel R I, Phillips R E

The diffusion of copper, manganese and zinc as affected by  
concentration, clay mineralogy and associated anions.

Soil.Sci.Soc.Amer.Proc 1970, 34,866

24. Mott G J B

Thesis, Univ Oxford 1967 (see p 241 in ref 15)

25. Calvet R

Ann Agron 1973, 24,77-217

# List of SKB reports

## Annual Reports

1977-78

TR 121

**KBS Technical Reports 1 – 120.**  
Summaries. Stockholm, May 1979.

1979

TR 79-28

**The KBS Annual Report 1979.**  
KBS Technical Reports 79-01 – 79-27.  
Summaries. Stockholm, March 1980.

1980

TR 80-26

**The KBS Annual Report 1980.**  
KBS Technical Reports 80-01 – 80-25.  
Summaries. Stockholm, March 1981.

1981

TR 81-17

**The KBS Annual Report 1981.**  
KBS Technical Reports 81-01 – 81-16.  
Summaries. Stockholm, April 1982.

1982

TR 82-28

**The KBS Annual Report 1982.**  
KBS Technical Reports 82-01 – 82-27.  
Summaries. Stockholm, July 1983.

1983

TR 83-77

**The KBS Annual Report 1983.**  
KBS Technical Reports 83-01 – 83-76.  
Summaries. Stockholm, June 1984.

1984

TR 85-01

**Annual Research and Development Report 1984**

Including Summaries of Technical Reports Issued during 1984. (Technical Reports 84-01-84-19)  
Stockholm June 1985.

1985

TR 85-20

**Annual Research and Development Report 1985**

Including Summaries of Technical Reports Issued during 1985. (Technical Reports 85-01-85-19)  
Stockholm May 1986.

1986

TR 86-31

**SKB Annual Report 1986**

Including Summaries of Technical Reports Issued during 1986  
Stockholm, May 1987

1987

TR 87-33

**SKB Annual Report 1987**

Including Summaries of Technical Reports Issued during 1987

Stockholm, May 1988

1988

TR 88-32

**SKB Annual Report 1988**

Including Summaries of Technical Reports Issued during 1988

Stockholm, May 1989

## Technical Reports

1989

TR 89-01

**Near-distance seismological monitoring of the Lansjärv neotectonic fault region Part II: 1988**

Rutger Wahlström, Sven-Olof Linder,  
Conny Holmqvist, Hans-Edy Mårtensson  
Seismological Department, Uppsala University,  
Uppsala  
January 1989

TR 89-02

**Description of background data in SKB database GEOTAB**

Ebbe Eriksson, Stefan Sehlstedt  
SGAB, Luleå  
February 1989

TR 89-03

**Characterization of the morphology, basement rock and tectonics in Sweden**

Kennert Röshoff  
August 1988

TR 89-04

**SKB WP-Cave Project  
Radionuclide release from the near-field in a WP-Cave repository**

Maria Lindgren, Kristina Skagius  
Kemakta Consultants Co, Stockholm  
April 1989

TR 89-05

**SKB WP-Cave Project  
Transport of escaping radionuclides from the WP-Cave repository to the biosphere**

Luis Moreno, Sue Arve, Ivars Neretnieks  
Royal Institute of Technology, Stockholm  
April 1989

TR 89-06

**SKB WP-Cave Project**  
**Individual radiation doses from nuclides contained in a WP-Cave repository for spent fuel**

Sture Nordlinder, Ulla Bergström  
Studsvik Nuclear, Studsvik  
April 1989

TR 89-11

**Prediction of hydraulic conductivity and conductive fracture frequency by multi-variate analysis of data from the Klipperås study site**

Jan-Erik Andersson<sup>1</sup>, Lennart Lindqvist<sup>2</sup>  
<sup>1</sup> Swedish Geological Co, Uppsala  
<sup>2</sup> EMX-system AB, Luleå  
February 1988

TR 89-07

**SKB WP-Cave Project**  
**Some Notes on Technical Issues**

Part 1: Temperature distribution in WP-Cave: when shafts are filled with sand/water mixtures  
Stefan Björklund, Lennart Josefson  
Division of Solid Mechanics, Chalmers University of Technology, Gothenburg, Sweden  
Part 2: Gas and water transport from WP-Cave repository  
Luis Moreno, Ivars Neretnieks  
Department of Chemical Engineering, Royal Institute of Technology, Stockholm, Sweden  
Part 3: Transport of escaping nuclides from the WP-Cave repository to the biosphere.  
Influence of the hydraulic cage  
Luis Moreno, Ivars Neretnieks  
Department of Chemical Engineering, Royal Institute of Technology, Stockholm, Sweden

August 1989

TR 89-12

**Hydraulic interference tests and tracer tests within the Brändan area, Finnsjön study site**  
**The Fracture Zone Project – Phase 3**

Jan-Erik Andersson, Lennart Ekman, Erik Gustafsson, Rune Nordqvist, Sven Tirén  
Swedish Geological Co, Division of Engineering Geology  
June 1988

TR 89-13

**Spent fuel**  
**Dissolution and oxidation**  
**An evaluation of literature data**

Bernd Grambow  
Hanh-Meitner-Institut, Berlin  
March 1989

TR 89-08

**SKB WP-Cave Project**  
**Thermally induced convective motion in groundwater in the near field of the WP-Cave after filling and closure**

Polydynamics Limited, Zürich  
April 1989

TR 89-14

**The SKB spent fuel corrosion program**  
**Status report 1988**

Lars O Werme<sup>1</sup>, Roy S Forsyth<sup>2</sup>  
<sup>1</sup> SKB, Stockholm  
<sup>2</sup> Studsvik AB, Nyköping  
May 1989

TR 89-09

**An evaluation of tracer tests performed at Studsvik**

Luis Moreno<sup>1</sup>, Ivars Neretnieks<sup>1</sup>, Ove Landström<sup>2</sup>  
<sup>1</sup> The Royal Institute of Technology, Department of Chemical Engineering, Stockholm  
<sup>2</sup> Studsvik Nuclear, Nyköping  
March 1989

TR 89-15

**Comparison between radar data and geophysical, geological and hydrological borehole parameters by multivariate analysis of data**

Serje Carlsten, Lennart Lindqvist, Olle Olsson  
Swedish Geological Company, Uppsala  
March 1989

TR 89-10

**Copper produced from powder by HIP to encapsulate nuclear fuel elements**

Lars B Ekbohm, Sven Bogegård  
Swedish National Defence Research Establishment  
Materials department, Stockholm  
February 1989

TR 89-16

**Swedish Hard Rock Laboratory – Evaluation of 1988 year pre-investigations and description of the target area, the island of Åspö**

Gunnar Gustafsson, Roy Stanfors, Peter Wikberg  
June 1989

TR 89-17

**Field instrumentation for hydrofracturing stress measurements  
Documentation of the 1000 m hydrofracturing unit at Luleå University of Technology**

Bjarni Bjarnason, Arne Torikka  
August 1989

TR 89-18

**Radar investigations at the Saltsjötunnel – predictions and validation**

Olle Olsson<sup>1</sup> and Kai Palmqvist<sup>2</sup>

<sup>1</sup> Abem AB, Uppsala, Sweden

<sup>2</sup> Bergab, Göteborg

June 1989

TR 89-19

**Characterization of fracture zone 2, Finnsjön study-site**

Editors: K. Ahlbom, J.A.T. Smellie, Swedish Geological Co, Uppsala

Part 1: Overview of the fracture zone project at Finnsjön, Sweden

K. Ahlbom and J.A.T. Smellie. Swedish Geological Company, Uppsala, Sweden.

Part 2: Geological setting and deformation history of a low angle fracture zone at Finnsjön, Sweden

Sven A. Tirén. Swedish Geological Company, Uppsala, Sweden.

Part 3: Hydraulic testing and modelling of a low-angle fracture zone at Finnsjön, Sweden  
J-E. Andersson<sup>1</sup>, L. Ekman<sup>1</sup>, R. Nordqvist<sup>1</sup> and A. Winberg<sup>2</sup>

<sup>1</sup> Swedish Geological Company, Uppsala, Sweden

<sup>2</sup> Swedish Geological Company, Göteborg, Sweden

Part 4: Groundwater flow conditions in a low angle fracture zone at Finnsjön, Sweden

E. Gustafsson and P. Andersson. Swedish Geological Company, Uppsala, Sweden

Part 5: Hydrochemical investigations at Finnsjön, Sweden

J.A.T. Smellie<sup>1</sup> and P. Wikberg<sup>2</sup>

<sup>1</sup> Swedish Geological Company, Uppsala, Sweden

<sup>2</sup> Swedish Nuclear Fuel and Waste Management Company, Stockholm, Sweden

Part 6: Effects of gas-lift pumping on hydraulic borehole conditions at Finnsjön, Sweden

J-E. Andersson, P. Andersson and E. Gustafsson. Swedish Geological Company, Uppsala, Sweden

August 1989

TR 89-20

**WP-Cave - Assessment of feasibility, safety and development potential**

Swedish Nuclear Fuel and Waste Management Company, Stockholm, Sweden

September 1989

TR 89-21

**Rock quality designation of the hydraulic properties in the near field of a final repository for spent nuclear fuel**

Hans Carlsson<sup>1</sup>, Leif Carlsson<sup>1</sup>, Roland Pusch<sup>2</sup>

<sup>1</sup> Swedish Geological Co, SGAB, Gothenburg, Sweden

<sup>2</sup> Clay Technology AB, Lund, Sweden

June 1989

TR 89-22

**Diffusion of Am, Pu, U, Np, Cs, I and Tc in compacted sand-bentonite mixture**

Department of Nuclear Chemistry, Chalmers University of Technology, Gothenburg, Sweden

August 1989

TR 89-23

**Deep ground water microbiology in Swedish granitic rock and its relevance for radionuclide migration from a Swedish high level nuclear waste repository**

Karsten Pedersen

University of Göteborg, Department of Marine microbiology, Gothenburg, Sweden

March 1989